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A temporal analysis of the consequences of the drought regime on the water footprint of agriculture in the Guadalupe Valley, Mexico

Vanessa Novoa^{1,6}, Carolina Rojas^{2,6}, Octavio Rojas³, Ramón Ahumada-Rudolph^{4✉} & Rebeca Moreno-Santoyo⁵

Changes in water availability have a substantial impact on the sustainability and maintenance of agriculture, with water footprint (WF) being a robust methodology to assess these transformations. The Guadalupe Valley is one of the places with the highest agricultural production in Mexico. Despite its semi-arid climatic conditions, it provides high-quality crops that are well-positioned in the world. The historical trend of rainfall and temperatures between 1987 and 2017 was analyzed to identify climatic patterns in the territory. Through the calculations of the water footprint of Grapevine and Olive crops, the sensitivity of the crops to recurrent water deficit and their adaptation in their yields to drought episodes was identified. The reduction in precipitation and occurrence of extreme temperatures have contributed significantly towards augmenting crop evapotranspiration and, consequently, intensifying crop irrigation demands. As a result, there has been an apparent increase in the consumption of $WF_{\text{agricultural}}$ since 2007. Thus, the period of highest $WF_{\text{agricultural}}$ consumption was 2014 (Extremely dry), as opposed to 2011 (Very wet). In particular, the lowest WF_{green} consumptions were observed in extremely dry years, that is, > 20% of the $WF_{\text{agricultural}}$ intensifying drought events. Therefore, these periods were compensated with higher uses of WF_{blue} and WF_{gray} , which are inversely correlated with precipitation, where vine crops consume 73% more $WF_{\text{agricultural}}$ compared to olive plantations, showing greater interannual variability. These results contribute to analyzing the temporal evolution of water consumption for agriculture, providing a basis for rational water use strategies.

Keywords Drylands, Arid, Semi-arid, Blue- Green- Grey water, Water scarcity

Abbreviations

AR (kg/hamass/area)

Application rate of a chemical (fertilizer or pesticide) per unit of land

α

Leaching-run-off fraction, i.e. fraction of applied Chemicals reaching freshwater bodies

CWU_{blue} ($\text{m}^3 \text{ ha}^{-1}/\text{volume}/\text{area}$)

Blue crop water use

CWU_{green} ($\text{m}^3 \text{ ha}^{-1}/\text{volume}/\text{area}$)

Green crop water use

¹Instituto de Alta Investigación, Universidad de Tarapacá, 18 de Septiembre 2222, Arica, Chile. ²Facultad de Arquitectura, Diseño y Estudios Urbanos, Instituto de Estudios Urbanos y Territoriales, Instituto Milenio de Socio-Ecología Costera SECOS, Centro de Desarrollo Urbano Sustentable CEDEUS, Pontificia Universidad Católica de Chile, El Comendador 1916, Providencia, Santiago, Chile. ³Departamento de Planificación Territorial y Sistemas Urbanos, Facultad de Ciencias Ambientales, Centro EULA, Universidad de Concepción, Víctor Lamas 1290, PO Box 160-C., Concepción, Chile. ⁴Laboratorio de Química Aplicada y Sustentable (LabQAS), Departamento de Química, Facultad de Ciencias, Universidad del Bío-Bío, Avenida Collao 1202, PO Box 5-C., 4051381 Concepción, Chile. ⁵Facultad de Ciencias Marinas, Instituto de Investigaciones Oceanológicas, Universidad Autónoma de Baja California, Carretera Ensenada-Tijuana 3917, Ensenada, Baja California, Mexico. ⁶These authors contributed equally: Vanessa Novoa and Carolina Rojas. ✉email: reahumada@ubiobio.cl

CWR (mm month ⁻¹ /length/time)	Crop water requirement
C _{max} (kg m ³ /mass/volume)	Maximum acceptable concentration of a chemical in a receiving water body
C _{nat} (kg m ³ /mass/volume)	Natural concentration of a chemical in the receiving water body
ET c(mm month ⁻¹ /length/time)	Crop evapotranspiration (under optimal conditions)
ET _{blue} (mm year ⁻¹ / mm month ⁻¹ /length/time)	Blue water evapotranspiration
ET _{green} (mm year ⁻¹ /mm month ⁻¹ /length/time)	Green water evapotranspiration
ETo mm (month ⁻¹ /length/time)	Reference crop evapotranspiration
I _{gp}	Length, days in each stage of the cycle
IR (mm month ⁻¹ /length/time)	Irrigation requirement
Kc	Crop coefficient
Pe _{ff} (mm month ⁻¹ /length/time)	Effective rainfall
WF _{agricultural} (m ³ t ⁻¹ /volume/time)	Agricultural water footprint
WF _{blue} (m ³ t ⁻¹ /volume/time)	Blue water footprint
WF _{crops} (m ³ t ⁻¹ /volume/time)	Crops water footprint
WF _{gray} (m ³ t ⁻¹ /volume/time)	Gray water footprint
WF _{green} (m ³ t ⁻¹ /volume/time)	Green water footprint
WF _{grapes} (m ³ t ⁻¹ /volume/time)	Grapes water footprint
WF _{olives} (m ³ t ⁻¹ /volume/time)	Olives water footprint

At present, droughts are one of the extreme socio-natural hazards that pose the greatest severity or threat to social stability and biophysical resilience of the environment, due to their repercussions on agricultural production and economic growth¹. Moreover, drought processes develop progressively within a given territory due to a continuous and cyclical period of water stress, resulting from the abnormal decrease in precipitation, together with fluctuating temperatures and eco-hydrological changes².

It is essential to consider that the water cycle in environments affected by drought is characterized by high evapotranspiration rates and reduced precipitation, resulting in deficient surface runoff and low groundwater recharge. The latter leads to decreased soil moisture storage, increased erosion, and deteriorated vegetation cover in areas far from river sources^{3–5}.

In addition, droughts can be categorized into meteorological, agricultural, hydrological, and socioeconomic types based on their periodicity. They often lead to water shortages or deficits, particularly severe in arid or semi-arid regions⁶. These regions constitute 41% of the global area and are experiencing trends in water availability that fall below historical averages⁷. It is projected that dryland extent will increase by 4% to 10% by 2100 compared to the 1961–1990 period under the RCP4.5 and RCP8.5 climate change^{8,9}.

Water scarcity is a significant issue in various parts of the world. The critical areas affected by it include the western USA, Mexico, the west coast of South America, sub-Saharan Africa, Pakistan, the Middle East, and Central America^{5,8}. Nevertheless, the highest percentage of drylands is concentrated in Latin American and Caribbean countries, that is, Argentina (69%), Mexico (65%), and Chile (58%)¹⁰.

Furthermore, the combination of climate change, climate variability, and inefficient water usage¹¹, along with irrational consumption habits¹², intensified agriculture, and changes in land use¹³, have significantly impacted the load capacity of water resources and their security. This has led to a severe water crisis in arid areas¹⁴. As a result, conflicts related to socio-environmental issues have become more severe in various regions, hindering the achievement of the United Nations' Sustainable Development Goals (SDGs). Specifically, it is making it difficult to achieve the targets of Zero Hunger (Target 2), Clean Water and Sanitation (Target 6), and Life on Land (Target 15) by 2030^{15,16}.

However, it is not only about the amount of water that is limited by climatic variables or overexploitation of water sources, but also about the repercussions in geographical and temporal dimensions that influence the hydro-social cycle¹⁷. As agriculture is one of the main productive activities globally, extreme droughts have had an impact on the loss of agricultural land, estimated at approximately 21 t⁻¹ ha⁻¹ average, which is up to 16 times higher than the weathering process of rocks¹⁴.

Understanding freshwater consumption in agriculture is crucial¹⁸ to establishing resilience and food security strategies. Therefore, broadening the concept towards an environmental water flow approach becomes necessary. This approach will help in identifying water-saving routes in regional crop production. The assessment of water resources appropriation needs to broaden the criteria for integrated resource management. The focus should be on determining the various environmental sources or flows of water for effective and sustainable management of water resources¹⁹.

It is essential to consider not only the use of blue water (water resources extracted from surface water runoff and groundwater infiltration) but also the use of green water (precipitation temporarily stored in the soil or intercepted by vegetation and exported to the atmosphere as evapotranspiration), along with their interactions²⁰. The distribution of seasonal precipitation, combined with evapotranspiration, directly influences water availability patterns, affecting regional hydrology and crop yields²¹.

Consequently, the main functions of green water are productive (i.e., maintenance of up to ~65% of overall crop production) and regulatory (i.e., energy balance and convective conditions)²², where green water flow is composed of about 59% transpiration, 21% plant interception, 10% soil interception, and 6% soil moisture evaporation²³.

Similarly, blue water serves a multitude of functions, including but not limited to supply, transport/chemical loading (nutrients and pollutants), regulation/control, and production^{24,25}; furthermore, irrigation accounts for about 70% of the world's blue water use²⁶.

A complex interaction of these variables (climatic and productive) has led water-scarce regions to produce more water-intensive but higher-value crops. In regions where water resources for agricultural purposes are scarce, a potential reduction of 56% in water usage can be achieved by cultivating crops better adapted to the area's green and blue water availability²⁷.

At the same time, it is essential to acknowledge that soils in arid regions are fragile (eroded or degraded). This is because they have less natural primary production, organic matter, nutrients, water retention capacity, and fertility. As a result, plant growth and biomass are negatively affected. Therefore, before adding excessive amounts of chemical fertilizers, it is essential to assess the volume of grey water needed to assimilate the pollutants generated during the cultivation process²⁸.

Considering the previously mentioned, prioritizing the evaluation of water abstractions is crucial for understanding the relationship between productive activities, the water cycle, and the pressure on water resources to ensure sustainable development in agriculture. Comprehensive indicators such as estimating the water footprint can help achieve this goal²⁰.

In addition, the agricultural water footprint ($WF_{\text{agriculture}}$) is a measure that indirectly evaluates the impact of aridity by considering the local water demands and supplies²³. It considers atmospheric water requirements such as evaporation and evapotranspiration, as well as temperature, precipitation, and irrigation. The measure also considers soil moisture, soil characteristics, and crop yield^{21,29}.

$WF_{\text{agriculture}}$ studies distinguish between the use of different water sources—blue (WF_{blue}), green (WF_{green}), and gray (WF_{gray})—where the amount of precipitation before the growing season can reduce the need for irrigation during crop growth. This can lead to a significant decrease in the use of blue water up to 96% for some crops^{8,20,30}.

However, during drought events, scarcity affects water flows equally, becoming crucial to increase irrigation to compensate for crops' deficit or green water requirements (CWU_{green}). Nevertheless, more than blue water requirements (CWU_{blue}) are required, constituting deficit irrigation at the expense of environmental flows and/or groundwater reserves^{30–32}.

Thus, to contribute to land management programs for irrigated crops in water-stressed basins, it is crucial to have accurate information on how agricultural water footprints have changed over time. This requires refined estimates over extended periods, which can be used in decision-making to reduce the latency of input data and improve future projections. By incorporating the influence of intra- and inter-annual climate variability in drought mitigation or action plans, we can make more informed decisions based on water footprint results³³.

Our research lies in understanding water use in agriculture, considering that in Mexico, approximately two-thirds of its territory is located in arid or semi-arid areas facing natural water scarcity, where the pressure on water resources has experienced a severe drought condition, which has persisted in the last 27 years, associated with El Niño Southern Oscillation (ENSO) and by the increase in water demand (up to 35%), especially after the mid-1990s^{34–36}.

The water deficit is particularly severe in northern Mexico, where semi-arid areas cover about 50% of the total surface³⁷. The central forecasts correspond to a 10–20% decrease in mean annual precipitation, a 1.5 to 2.5 °C temperature increase, intensified incidence of heat waves, and a reduction in aquifer recharge³⁸.

The Valle de Guadalupe is a semi-arid rural area in the Ensenada municipality in Baja California. Despite the critical water situation, it is one of Mexico's highest producers of olives and grapevines, with about 46.6% of the total cultivated area dedicated to high-quality, high-value grapes³⁹. The area produces 75% of Mexico's wine and 30% of its olive oil, making export crops and by-products the primary economic and tourism livelihood source. However, the limited water resources and agricultural land make the continuity of these industries vulnerable.

Different crops require varying amounts of water, and it is crucial to assess their water footprint to understand the factors that influence these differences. This helps promote the efficient use of water in agricultural production. To this end, this study examines the water footprint for olive and vine cultivation, two of the most predominant crops. Additionally, the impact of climate variability on agricultural $WF_{\text{agricultural}}$ consumption is analyzed through temporal analysis. This analysis is based on the climatic variables recorded for very wet, wet, normal, dry, and arid years.

Results

Determination of climatic variability

The interannual precipitation of the Guadalupe Valley for the 29 years analyzed shows a historical mean of 250 mm, with 62% of the analyzed years below this mean. The years 1989, 1999, 2002, 2007, 2009, and 2014 recorded amounts between 59–133 mm, which is significantly lower precipitation compared to the other years ($F_{(30,341)} = 1.03$; $p = 0.001$) (Fig. 1a). That is, years with a deficit between 76 to 47% of precipitation regarding the 1987–2017 average, with a negative trend starting in 2005 (Fig. 1a).

In the annual cycle, the behavior of seasonal cycles and extreme values in daily and monthly observations indicate greater variability and rainfall amounts during the winter period (December–March), where about 74% of precipitation occurs ($ANOVA F_{(11,360)} = 16.8$, $p = 0.001$), with a prolonged dry period in June–August (Fig. 1b).

According to the percentile analysis, extreme conditions of very dry years were observed in 1989, 1999, 2002, 2007, 2009, and 2014 with 0–20% of the precipitation percentile, with 1989 being the most extreme. Contrary to the years 1987, 1992, 1993, 1998, 1998, 2004, 2010, and 2011 which showed a very wet condition, between 80–100%. Only five years of the analyzed series were classified as normal (Fig. 2).

Regarding temperature, annual mean values ranged from 15° to 19.5 °C, with the years 2006 and 2014 being the warmest with recorded temperatures above 19 °C, that is, positive anomalies of + 1.4 °C to + 1.8 °C above the

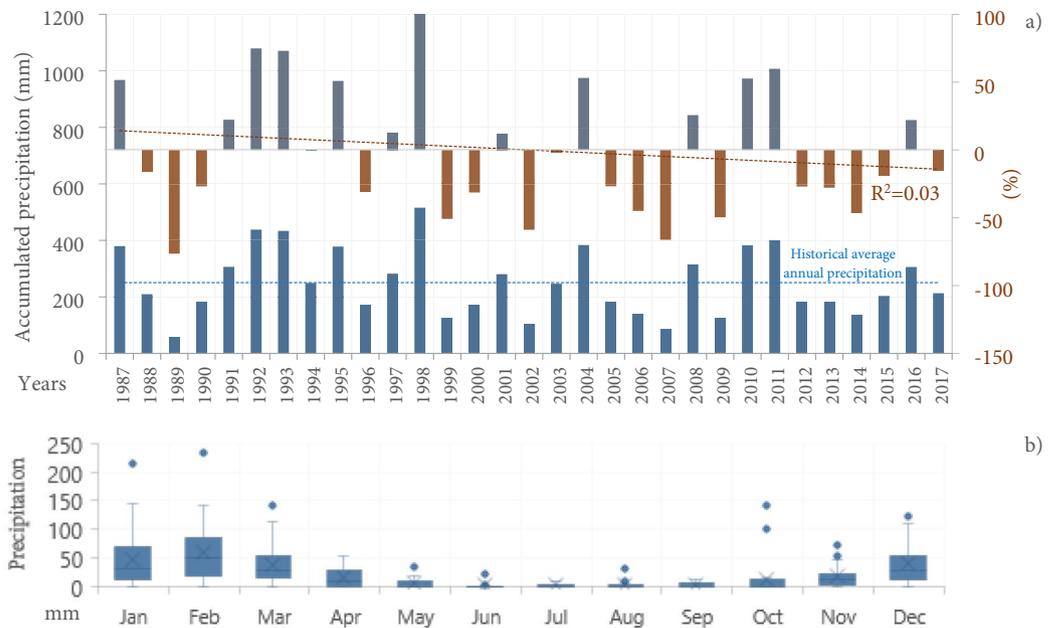


Figure 1. Precipitation behavior from 1987—2017 of the Guadalupe Valley. **(a)** Percentage anomaly of total annual precipitation: blue bars represent positive values (surplus), and red bars represent negative values (deficit) and accumulated frequency of annual precipitation in both climatological stations Agua Caliente and El Porvenir, **(b)** Trends of monthly periods (boxplots): the horizontal line and square inside the box indicate the median and mean. The lower and upper ends of the box correspond to the quantiles 0.25 and 0.75. Lower and upper whiskers: quantile 0.05 and 0.95, respectively.

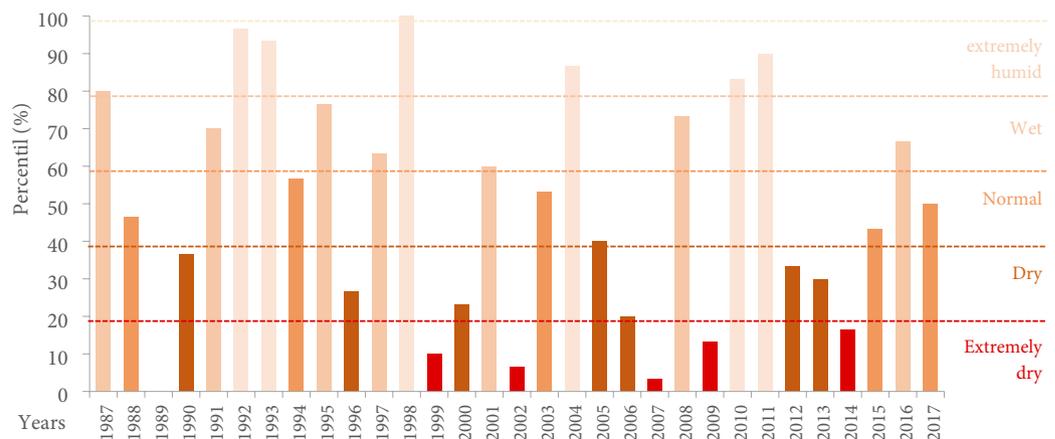


Figure 2. Climate classification according to the precipitation thresholds in the Guadalupe Valley.

climatological average in 1987–2017 series (30 years), where it is possible to observe a dominance of the natural climate variability signal, as opposed to ENSO years (contributing to warmer temperatures, overlapping with the El Niño period 2005–2006 or 2014–2015) (Fig. 3a).

The evolution of the monthly mean temperature shows significantly higher values in the summer period (July, August, September) compared to other months of the year (ANOVA $F_{(11,360)} = 170.2$, $P = 0.001$) (Fig. 3b).

Lastly, the precipitation and temperature data series are moderately inversely correlated ($r = -0.66$; $p < 0.05$), i.e., precipitation is lower in the months where temperature increases, decreasing water availability in the Guadalupe Valley in the warmer periods.

In summary, high climatic variability of the Guadalupe Valley can be observed from 1996, increasing both ‘very dry’ and ‘very wet’ years, thus intensifying extreme weather conditions. As a result, the occurrence of extreme meteorological droughts has been evident every 3 to 5 years since 1999.

The annual precipitation trend has decreased over the last 12 years; however, 1989 stands out for its low rainfall and high temperatures. Meanwhile, the changes identified in the annual temperature result in an amplification of the thermal oscillation.

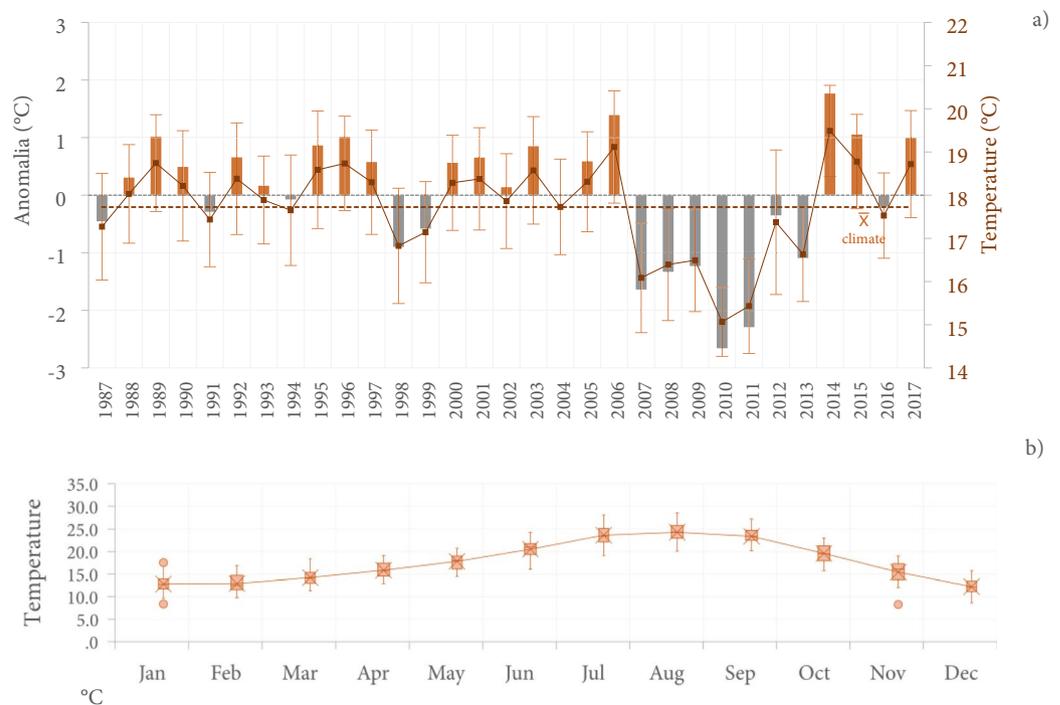


Figure 3. Evolution of the average temperature of the Guadalupe Valley. **(a)** Absolute anomalies of the average annual temperature for the 1987–2017 period (red bars indicate warming, and blue bars indicate cooling) and annual average, **(b)** monthly average.

Water footprint of the Guadalupe Valley for the agricultural sector

Crop evapotranspiration (ETc) in 2007–2017 was estimated between 573 and 830 mm. It is observed that ETc values decreased in 2010–2011, i.e., 596–573 mm respectively, unlike 2014, where it reached the maximum amounts. The temporal behavior of ETc is correlated with ETo ($r = 0.99$; $p < 0.05$) and irrigation requirements (IR) ($R = 0.93$; $p < 0.05$) (Fig. 4).

Similarly, the IR of the crops is inversely correlated with precipitation and effective precipitation ($r = -0.84$; $p < 0.05$), which were higher in 2010–2011. With the increase in precipitation, ETo and ETc values were drastically reduced, combined with the decrease in temperatures, as in the case of 2010–2011. In contrast, in 2014, although it was not the only extremely dry year with a precipitation deficit, the extreme temperatures recorded led to an increase in ETc and IR.

The $WF_{\text{agricultural}}$ of the crops analyzed in the Guadalupe Valley presented estimated consumptions between 645 and 1121 m^3/ton . Although there was some interannual variability, a slight incremental trend was identified in the 11 years analyzed, and the year 2014, which had extremely dry rainfall, had the highest $WF_{\text{agricultural}}$ consumption. On the other hand, 2011, which was a very wet year, had the lowest consumption estimated (as shown in Fig. 4).

Thus, $WF_{\text{agricultural}}$ is significantly correlated with ETo ($r = 0.86$; $p < 0.05$), ETc ($r = 0.88$; $p < 0.05$), IR ($r = 0.90$; $p < 0.05$), WF_{blue} ($r = 0.95$; $p < 0.05$), and WF_{gray} ($r = 0.93$; $p < 0.05$), and inversely correlated with precipitation ($r = -0.85$; $p < 0.05$) (Fig. 4).

The highest consumption of WF_{green} was observed in 2010, 2011, and 2015 (Extremely humid and normal), with 35%, 38%, and 27% of the total $WF_{\text{agricultural}}$ being found in these periods the precipitation and soil moisture necessary for crop growth, a variable that is inversely correlated with IR ($r = -0.78$; $p < 0.05$).

While the highest uses of WF_{blue} were calculated for Extremely dry years, when the need for irrigation increased, that is, the years 2007 ($620 m^3 t^{-1}$), 2009 ($571 m^3 t^{-1}$), 2014 ($752 m^3 t^{-1}$) and 2017 ($620 m^3 t^{-1}$), volumes corresponding to 64–68% of the $WF_{\text{agricultural}}$. Additionally, WF_{blue} trends are correlated with ETo ($r = 0.81$; $p < 0.05$), ETc ($r = 0.80$; $p < 0.05$), IR ($r = 0.97$; $p < 0.05$), and WF_{gray} ($r = 0.95$; $p < 0.05$) and inversely correlated with precipitation ($r = -0.87$; $p < 0.05$).

Meanwhile, the highest uses of WF_{gray} were determined in 2007, 2009, 2014, and 2015 (Extremely dry and normal) with consumption above $200 m^3 t^{-1}$, a variable correlated with irrigation requirements ($r = 0.82$; $p < 0.05$), and inversely correlated with precipitation ($r = -0.91$; $p < 0.05$) (Fig. 4).

Water footprint by crop type

Significant differences were observed in the $WF_{\text{agricultural}}$ consumption of the evaluated crops, with higher consumption in grapevine plantations by 73%, corresponding to an average of $127.4 \pm 11.8 m^3 t^{-1}$ between 2007–2017 ($F_{(1,20)} = 185.1$; $p = 0.001$), as well as the consumptions of WF_{green} , WF_{blue} , and WF_{gray} , which are significantly

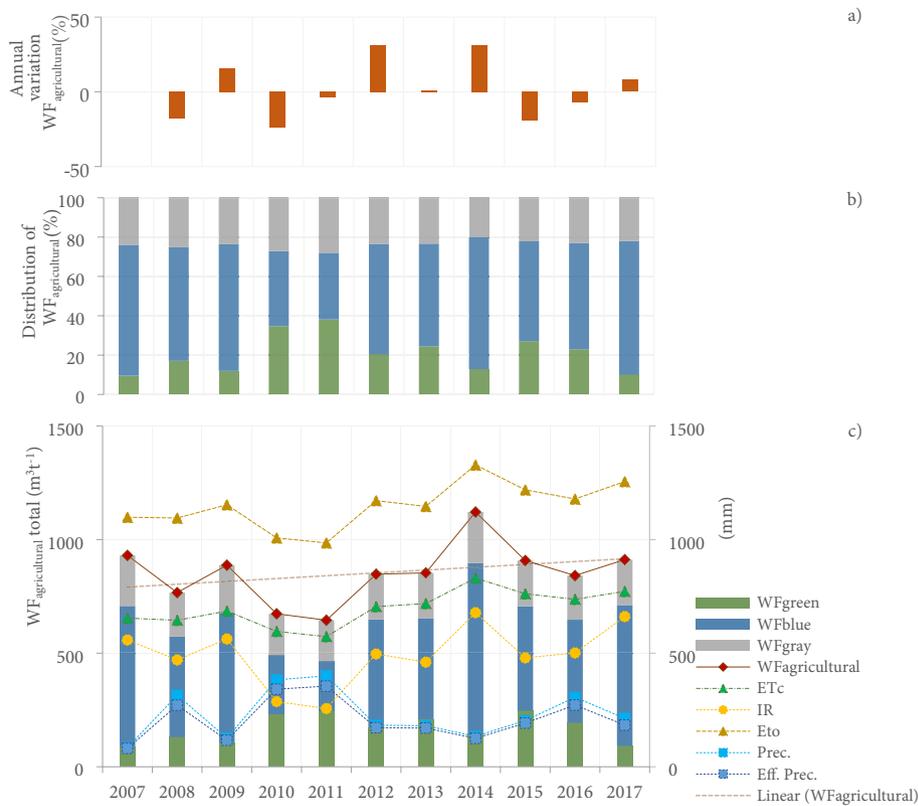


Figure 4. (a) Annual variation percentage of $WF_{agricultural}$ regarding the previous year. (b) Annual distribution in WF_{blue} , WF_{green} , and WF_{gray} consumptions. (c) Accumulated $WF_{agricultural}$ of the Guadalupe Valley basin, Mexico, crop evapotranspiration (ETc), irrigation requirements (IR), reference evapotranspiration (Eto), annual precipitation (Prec.) and effective precipitation (Eff. Prec.) from 2007–2017.

higher in grape production compared to olive, that is, 67%, 84%, and 43%, respectively ($F_{(3,30)} = 129.7$; $p = 0.001$) (Fig. 5).

Grapevine crops show high interannual variability, that is, an average of $3\% \pm 22.5$ in $WF_{agricultural}$ consumption, with a maximum of $922.1 \text{ m}^3\text{t}^{-1}$ in 2014 (Extremely dry), presenting a higher use of WF_{green} in 2010, 2011, and 2015, which is above the $170 \text{ m}^3\text{t}^{-1}$ uses of WF_{blue} in 2007, 2014, and 2017, exceeding $500 \text{ m}^3\text{t}^{-1}$, and WF_{gray} in 2007–2014, exceeding $150 \text{ m}^3\text{t}^{-1}$. Maximum water uses of CWU_{green} crops were evinced in the 2011 period (Extremely humid) ($1883 \text{ m}^3\text{ha}^{-1}$) and for CWU_{blue} in 2014 (Extremely dry) ($4601 \text{ m}^3\text{ha}^{-1}$) (Fig. 6a).

On the other hand, olive crops presented a lower interannual variability of $WF_{agricultural}$, that is $-1\% \pm 5.3$, with amounts up to $199 \text{ m}^3\text{t}^{-1}$ recorded in 2014 (Extremely dry). The highest consumptions of WF_{green} were in 2010, 2011, and 2013, above $55 \text{ m}^3\text{t}^{-1}$; WF_{blue} in 2014–2017 with values above $95 \text{ m}^3\text{t}^{-1}$; and WF_{gray} in 2016 with more

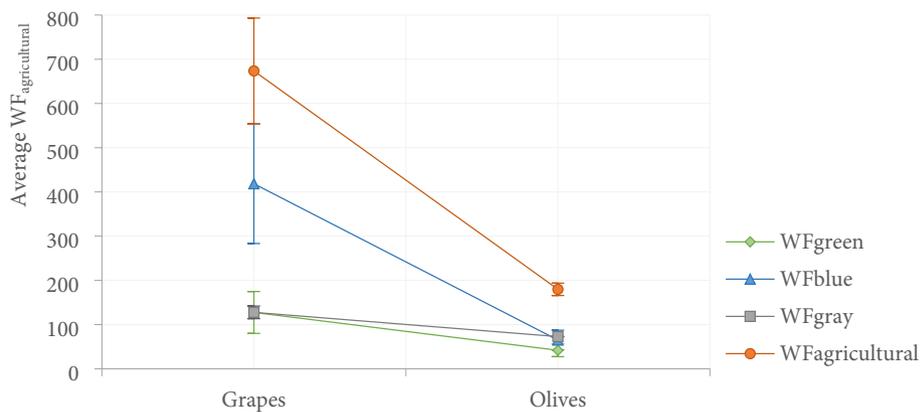


Figure 5. Average of $WF_{agricultural}$ (WF_{green} , WF_{blue} , WF_{gray}) the grapes and olives crops, series analyzed for the Valle de Guadalupe, Mexico.

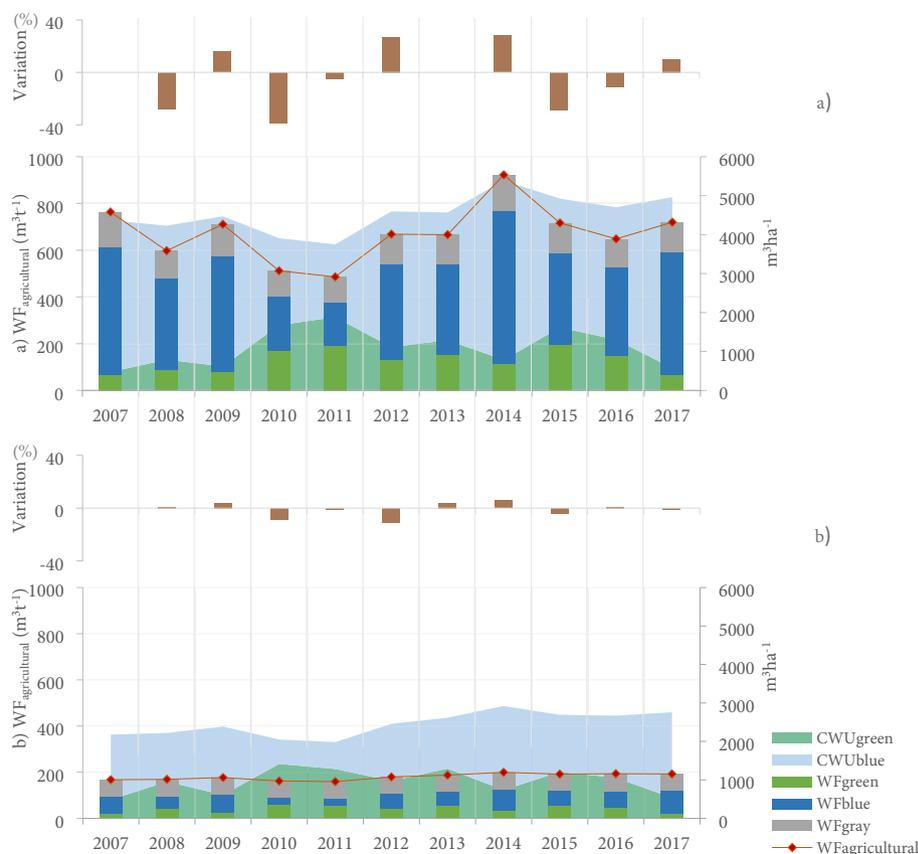


Figure 6. Annual variation of $WF_{\text{agricultural}}$ consumption of crops: a) grapes b) olive analyzed in 2007–2017 for the Valle de Guadalupe, Mexico.

than $74 \text{ m}^3 \text{ t}^{-1}$. In addition, in the 2010 period, the maximum uses of CWU_{green} ($1414 \text{ m}^3 \text{ ha}^{-1}$) and 2014–2017 of CWU_{blue} ($> 2000 \text{ m}^3 \text{ ha}^{-1}$) were estimated (Fig. 6b).

Crop conditions in the Guadalupe Valley reported yields below $10 \text{ t}^{-1} \text{ ha}^{-1}$ for grapevine^{39,40} and $12 \text{ t}^{-1} \text{ ha}^{-1}$ for olive⁴¹. Therefore, the net crop production and annual $WF_{\text{agricultural}}$ consumptions, when compared to the amount of water granted ($25 \times 10^6 \text{ m}^3$ according to the Drinking Water Committee), it is possible to identify the annual use of 36.4 to 82.7% of the available water (Table 1), where the maximum Total WF_{grapes} were evinced in 2012, 2013, 2014 and for Total WF_{olives} in 2010, 2013, 2016 (Table 1). Both crops accounted for an average of $55\% \pm 12$ of the water available for agriculture in the study period.

Years	Grapes production (t)	Total WF_{grapes} (m^3)	Olive production (t)	Total WF_{olives} (m^3)	Total $WF_{\text{grapes} + \text{olives}}$ (m^3)	Fraction used (%)
2007	16,264.5	12,413,936.5	1968.6	329,074.6	12,743,011.1	51.0
2008	14,658.8	8,761,143.6	2009.3	339,631.3	9,100,774.8	36.4
2009	19,099.9	13,574,360.8	1814.0	320,602.7	13,894,963.5	55.6
2010	22,514.3	11,524,383.7	7196.8	1,162,996.2	12,687,379.8	50.7
2011	21,820.4	10,602,768.1	260.0	41,424.1	10,644,192.1	42.6
2012	23,146.0	15,470,895.8	1985.1	355,914.4	15,826,810.2	63.3
2013	22,349.4	14,890,058.8	5811.0	1,087,748.4	15,977,807.2	63.9
2014	22,411.2	20,664,300.5	0.0	0.0	20,664,300.5	82.7
2015	19,357.2	13,868,140.3	160.4	30,711.8	13,898,852.1	55.6
2016	15,919.6	10,319,803.6	4851.3	938,202.3	11,258,005.9	45.0
2017	20,678.5	14,875,620.4	2329.5	448,248.6	15,323,869.1	61.3

Table 1. WF_{grapes} and WF_{olives} consumption according to annual production and water availability in the Guadalupe Valley, Mexico. Source: SIAP, IMIP.

Discussion

Currently, drought management comprises measures aimed at restricting water supply, reducing demand, and minimizing impacts. These measures have multiple effects on the maintenance and conservation of ecosystems, optimization of water supply, availability, and recharge within watersheds⁴.

In order to use resources efficiently, it is important to understand the consequences of climate variability on hydrological responses. In terms of precipitation in the Guadalupe Valley, in the last 30 years analyzed, 62% of the period is below the historical average of 250 mm, i.e., a deficit of 76–47% of precipitation, with a clear negative annual/seasonal trend since 2005.

The yearly variation in precipitation levels can be attributed to the impact of El Niño-Southern Oscillation (ENSO) and Pacific Warm Decadal Oscillation (PDO) events⁴². Our analysis has identified extreme conditions such as very dry or wet years and their alternation as evidence of these events.

Therefore, it is important to note that the transition from El Niño during 1998 (the rainiest year in the hydro-climatic record, classified as very wet) to the La Niña event (the following year) (Fig. 1), which marked the beginning of prolonged drought periods, in which extreme drought became even more severe later, that is, in the years 1999, 2002, 2006, 2006, 2007, 2009, and 2014 ($P < 140$ mm), being possible to observe a dominance of the natural climate variability signal, versus ENSO years, contributing to warmer temperatures as El Niño in the 2005–2006 or 2014–2015 periods observed, with positive anomalies from +1.4 °C to +1.8 °C.

El Niño is linked to the increase in surface water temperature in the central Pacific Ocean, inducing severe droughts in Indonesia, Australia, and northeastern South America^{5,43}. La Niña, on the other hand, occurs when the surface water temperature drops off the coast in the Pacific Ocean, contributing to drier than normal weather in the North and South American regions, which also influences drought episodes, with a more complex impact on weather patterns than El Niño^{5,43,44}.

Although El Niño-Southern Oscillation (ENSO), Pacific Decadal Oscillation (PDO) and Sea Surface Temperatures (SST) of the tropical Atlantic have played an important role in recent drought in northern Mexico^{44,45}. However, global warming has increased the demand for atmospheric humidity and has probably altered circulation patterns, contributing to sustained episodes of drought in the northwestern region of Mexico, observed since 1994³⁴, where recurrent dry years have significantly diminished water resources, in which high interannual climate variability through changes in precipitation and temperature, aggravating the present situation in the Guadalupe Valley¹⁰.

Temperature and precipitation are the main climatic factors that produce changes in soil water content and, therefore, directly regulate crop growth. By including these variables, the water footprint analysis reflects the water requirements in agriculture and the sensitivity of the results, where the precipitation deficit and extreme increase in temperatures conditioned the increase in ETo, ETc and, IR⁴⁶.

Therefore, of the 11 years evaluated and considered as the present conditions (2007–2017), a slight incremental trend of water consumption was observed, despite the interannual variability, where the highest consumption of $WF_{\text{agricultural}}$ (grapevine + olive) was in the year 2014, classified as Extremely dry. This trend was also identified in the high WF_{blue} consumption in 2007, 2009, 2014, and 2017, since during drought years, there was an increase in crop ET related to climatic variables and irrigation requirements.

In contrast, in the years classified as extremely wet and normal, i.e., 2010, 2011, and 2015, more WF_{green} was consumed; being essential to include the evapotranspiration flux link between the ecological and hydrological systems, understanding that biological water use is inexorably linked to ecosystem productivity (Fig. 7).

The reported results, as described in Marston and Konar²⁵ in the California Valley, differences in $WF_{\text{agricultural}}$ during periods of drought are also attributed to reasons for changes in crop area and yields, increases in ETc, irrigation demand to compensate for precipitation deficits, and increased groundwater consumption due to reduced surface water availability⁴⁷.

In these wine and fruit growing regions, in addition to the seasonal climatic changes described above, it is also necessary to include the influence of climate on an intermediate geographical scale (< 50 km) since these local climatic differences can cause considerable changes in crop physiology and phenology. Therefore, it is suggested to grow deep-rooted plants in drylands, due to the lower probability of suffering stress due to effects or changes in soil moisture, compared to shallow-rooted crops with less root coverage^{48,49}.

However, crops can also modify their water needs through physiological adjustments, that is, in the regulation of stomatal conductance and leaf area transformation⁵⁰, being crucial to the adaptation of plants to changing hydrological regimes, where their growth depends on whether the altered water supply is sufficient. Thus, the highest water requirements per hectare of crops were recorded in 2011 (1883 m³ha⁻¹) for CWU_{green} and 2014 (4601 m³ha⁻¹) for CWU_{blue} , with olive crops being the best adapted to interannual variability according to $WF_{\text{agricultural}}$.

As $WF_{\text{agricultural}}$ is associated with both yield and seasonal water consumption, grapevine growing conditions in the Guadalupe Valley are limited by low yields, i.e., less than 10 t⁻¹ ha⁻¹, compared to the main wine producers, for example, Italy, France, and Spain (over 80 t⁻¹ ha⁻¹). For olives, yields are even more extreme, reaching maximums of 23.40 t⁻¹ ha⁻¹ and minimums of 0.10 t⁻¹ ha⁻¹, respectively^{35,37}.

Excessive use of nitrogen fertilizers is common in sandy soils such as those in the Guadalupe Valley. As a result, the soil contains nitrate levels of more than 15 mg kg⁻¹, while nearby aquifers show levels of 10 mg L⁻¹⁵¹. This has led to severe contamination problems by leaching and accumulation of macronutrients, especially during extremely dry periods when more irrigation is needed. As a solution, significant uses of WF_{gray} were established to prevent further contamination⁵².

For agriculture to be a sustainable activity over time, it is necessary to include all approaches for mitigation and adaptation to water-scarce environments, from broadly tolerant crops to the introduction of climate and disease-resistant seed varieties, diversification, and crop calendars. In addition to the above, better information

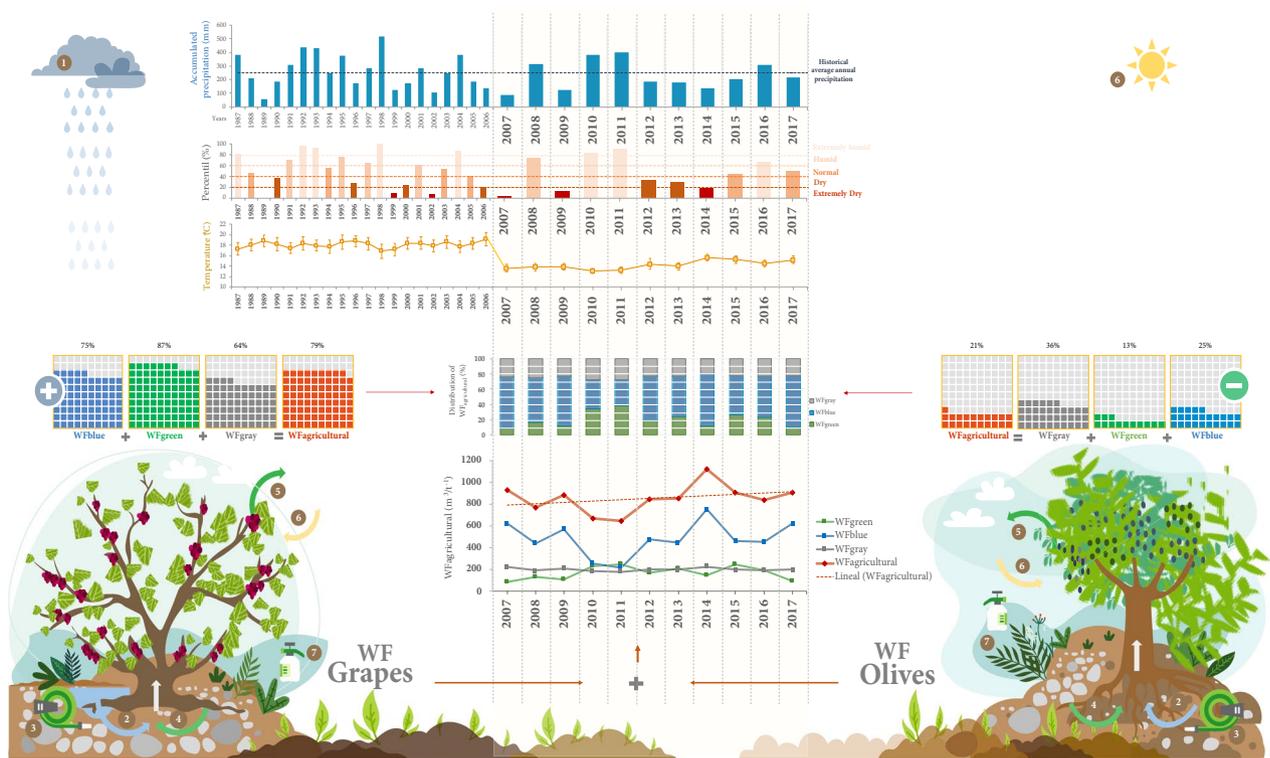


Figure 7. Summary of the agricultural water footprint of the Guadalupe Valley (Reference: Author’s elaboration).

on droughts and climate variables, planned and comprehensive agricultural relocation, support for land use changes, and marketing strategies⁵³ become necessary, with productive land and fertile soil being the most important natural capital asset²⁷. Therefore, soil preparation is essential, such as zero tillage and composting, the introduction of cover crops, intensive grazing management, and reduction in the use of chemical and synthetic inputs. Banwell⁵⁴ insist that such measures should consider the participation of various stakeholders, especially with a bottom-up approach.

Indeed, only by modifying the cropping pattern could water scarcity be significantly reduced and production increased. According to Davis⁵⁶, it is possible to reduce green and blue water consumption by 13.6% and 12.1%, respectively, only by growing the crops recommended for the area. First, however, farmers and decision-makers must be aware of the problem.

Since drought or water scarcity are dynamic, multiple indicators for their monitoring have been established⁵⁷. Although initially, the water footprint is a tool for the evaluation of water consumption in production chains, in this study, it is possible to establish the sensitivity of crop water consumption results to changing weather conditions^{58–60}, contributing to be the starting point for timely monitoring of water deficit and agricultural development in the area, for adequate preparation of responses according to the effects on the environment, agriculture, and local economy, understanding that the beginning and end of dry periods are perceived in retrospect. However, food security and water availability need immediate action^{48,61}.

Lastly, we must not forget that the water resource is a priority requirement on which we depend, fulfilling its function through its cycle, storage, transport, hydroclimatic and ecological regulation, in which water flows, in all its states (blue, green, and gray), intervene in energy balances and biogeochemical cycles, for example, critical ecosystem services require 90% of global evapotranspiration for their functioning^{62,63}.

Conclusion

The increase in precipitation in the Guadalupe Valley has a significant impact on the drastic reduction in ETO and ETC values, also influenced by the decrease in temperatures. However, this situation was contrary to the precipitation deficit and extreme temperatures that caused an increase in ETC and IR in crops.

A clear incremental trend in the consumption of $WF_{\text{agricultural}}$ was identified during the 11 years analyzed. Moreover, 2014 (Extremely dry) was the period with the highest $WF_{\text{agricultural}}$ consumption, as opposed to 2011 (very wet), where the lowest consumption was estimated.

The quantification of green water flow must necessarily be included in the evaluation of water resource security since the lowest consumption of WF_{green} was observed in extremely dry years, that is, > 20% of the $WF_{\text{agricultural}}$, due to the limitation in the availability of green water, intensifying and aggravating drought events. Periods compensated with higher uses of WF_{blue} , corresponding to 64–68% of the $WF_{\text{agricultural}}$, periods where the need for irrigation increases.

The dissolution of fertilizers (nitrogen) decreases in periods of water scarcity, thus increasing the probability of contamination of environmental matrices, where the highest WF_{gray} was determined in Extremely dry years with a consumption of over 200 m³/ton, a variable correlated with the irrigation requirement and inversely correlated with precipitation.

Grapevine crops consume 73% more $WF_{agricultural}$ compared to olive plantations and also showed greater inter-annual variability, in other words, less adaptation to climate variability. Nevertheless, the commercialization and water productivity of these crops is responsible for the continuity of their production.

The results obtained are a significant step towards reducing water waste in agriculture in arid regions with limited resources and fragile ecological environments. Despite the efforts made so far, such as the National Program Against Drought, the Mexican Drought Monitor, and the Programs for Drought Prevention and Mitigation Measures, these measures are considered insufficient. We must formulate strategies involving the rational use of agricultural water to tackle this issue effectively.

Materials and methods

Site of study

The Guadalupe Valley basin (32°03'N—116°37'W) is located northeast of the municipality of Ensenada, Baja California, Mexico (2380 km²). Its population reaches 6,924 inhabitants, concentrated in 0.4% of the basin's territory, 38.7% of the inhabitants live in poverty, and 35.5% are vulnerable due to social deprivation. Economic activities are mainly related to livestock, tourism, and agriculture⁶⁴.

Similarly, the little of the land in this area is used for agriculture (cultivated lands), accounting for 2.7% of the total land area (6272 hectares). Out of this, 60% is dedicated to grapevine and olive production. Other land uses include scrub (94.3%), forest (1.8%), grassland (1%) and water bodies (0.02%)^{64,65} as seen in Fig. 8.

The Guadalupe Valley is classified as a semi-desert region with a semi-arid Mediterranean climate, directly influenced by the Pacific Ocean coastal system, which is only 25 km away, exposing Guadalupe Valley to dense fogs, breezes, and marine humidity. Climatic conditions resulted in mild temperatures year-round, with an average of 18 °C annually from 1986–2016. Monthly temperatures range from 5.4 to 31.6°C⁶⁷.

Precipitation is a consequence of the weakening of the North Pacific system in the winter, which allows the entry of convective and frontal systems; average annual rainfall amounts reach 298 mm (1992–2016)⁶⁸. Precipitation variations are mainly associated with El Niño/Southern Oscillation and the Pacific Decadal Oscillation (PDO). These hydroclimatic phenomena bring unusual precipitation events to the region, as under warm conditions during El Niño and the PDO, the precipitation rate tends to increase. In contrast, under cold conditions during El Niño and the PDO, the precipitation decreases^{45,69}. The research methodology was developed in 3 stages. In the first stage, the climatic behavior of the Guadalupe Valley basin was analyzed. During the second

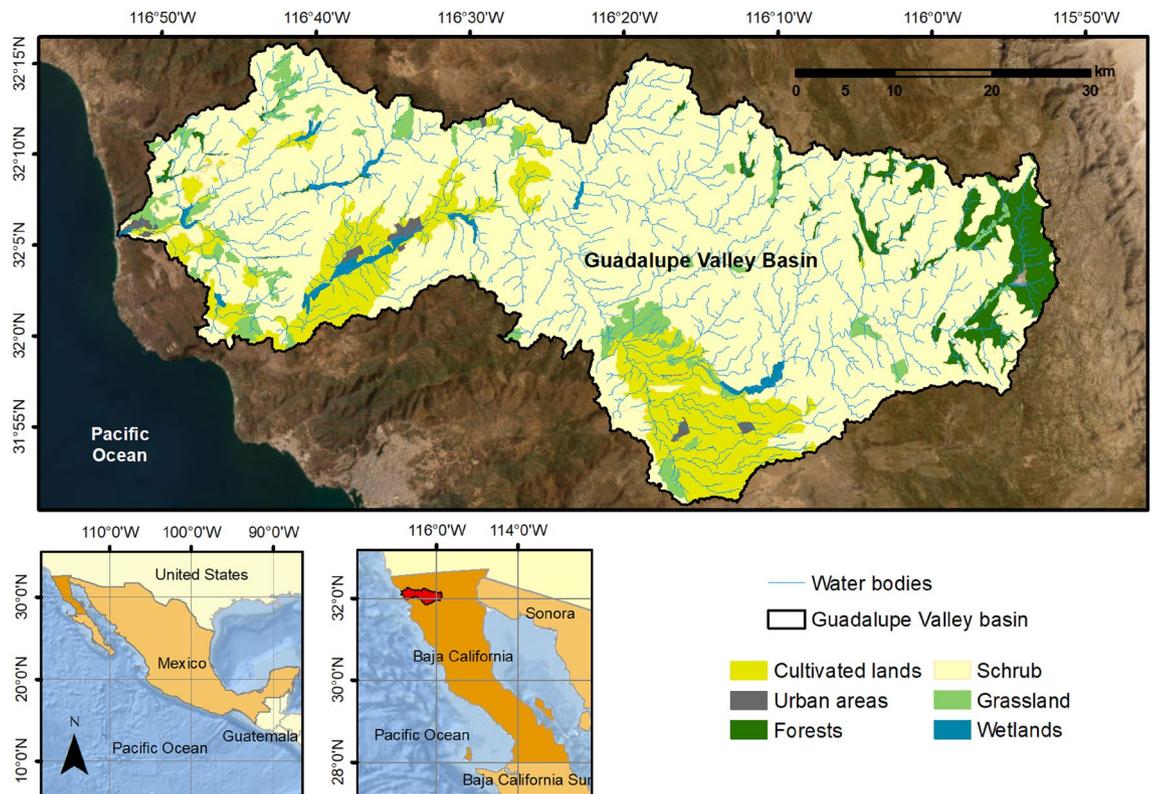


Figure 8. Location, land use, and water bodies identified in the study area, Guadalupe Valley, Mexico (Reference: ArcGIS 9.3 software developed by ESRI (Environmental Systems Research Institute), <https://www.esri.com>)⁶⁶.

stage, the database was developed, and the classification of the drought cycles was conducted. Finally, the crops' water footprint was calculated.

Determination of climatic variability

The meteorological input data were selected from 2 meteorological and pluviometric stations maintained by the Information System for Irrigation Water Management of Baja California (Fomento Agropecuario de Baja California) and the meteorological network CICESE-CLICOM (Center for Scientific Research and Higher Education at Ensenada). These stations are located in the Guadalupe Valley, that is Agua Caliente (32°04'N–116°37'W) and Porvenir (32°06'N–116°27'W) stations, considering a 31-year data (1987–2017 period). Monthly, seasonal, and annual precipitation and temperature data were used.

Exploratory and consistency analyses of the data were carried out using the double-mass curve method⁷⁰. Boxplots were used to visualize the behavior, which provided information on the trend of seasonal cycles and extreme values in the observations. The classification of dry, normal, and wet years was made using the percentile method, which defines five categories: very dry (0–20 percentile), dry (20–40 percentile), normal (40–60 percentile), wet (60–80) and very wet (80–100)⁷¹.

Calculation of the water footprint of olive and grapevine crops ($WF_{\text{agricultural}}$)

In the $WF_{\text{agricultural}}$ analysis, climate variables were included with georeferenced information from 2007–2017, with primary values of precipitation (mm/month), temperature (°C), wind (km/day), insolation (hrs), humidity (%), and secondary values of effective precipitation (mm/month), evapotranspiration (mm/day), solar radiation (MJ/m²/day), as well as standardized soil properties data.

In addition, grapevine and olive crop data were used, such as Kc values, stages, and critical depletion fraction by development period, along with estimates of crop evapotranspiration and irrigation requirements.

To estimate the $WF_{\text{agricultural}}$, the methodology proposed by Hoekstra⁷² was used, which considers the sum of its components: WF_{blue} , WF_{green} , and WF_{gray} of the assessed crops grapes (WF_{grapes}) and olives (WF_{olives}) (Eqs. (1) and (2)).

The determination of WF_{blue} and WF_{green} was made from Eqs. (3) and (4), where Y corresponds to crop yield (t ha⁻¹) and CWU_{blue} , CWU_{green} green and blue crop water use, according to its source (precipitation or irrigation), expressed in m³ ha⁻¹.

For the calculation of blue and green crop water use (CWU_{blue} , CWU_{green}), they were estimated based on the relationships of Eqs. (5) and (6). In the equations, Σ represents the crop growth cycle, that is, which starts from sowing (day 1) and ends at harvest. In addition, Igp represents the length in days of each cycle stage. ET_{blue} is blue water evapotranspiration, and ET_{green} is green water evapotranspiration. We use a conversion factor 10 to convert the water depths in millimeters into water volumes per land surface area in m³/ha.

$$WF_{\text{agricultural}} = \Sigma WF_{\text{crops}} (WF_{\text{grapes}} + WF_{\text{olives}}) \text{ m}^3 \text{ t}^{-1} \quad (1)$$

$$WF_{\text{crops}} = \Sigma WF_{\text{green}} + WF_{\text{blue}} + WF_{\text{gray}} \text{ m}^3 \text{ t}^{-1} \quad (2)$$

$$WF_{\text{green}} = CWU_{\text{green}} Y \text{ m}^3 \text{ t}^{-1} \quad (3)$$

$$WF_{\text{blue}} = CWU_{\text{blue}} Y \text{ m}^3 \text{ t}^{-1} \quad (4)$$

$$CWU_{\text{blue}} = 10 \times \sum_{d=1}^{Igp} ET_{\text{blue}} \text{ m}^3 \text{ ha}^{-1} \quad (5)$$

$$CWU_{\text{green}} = 10 \times \sum_{d=1}^{Igp} ET_{\text{green}} \text{ m}^3 \text{ ha}^{-1} \quad (6)$$

$$ET_{\text{blue}} = IR \text{ mm year}^{-1} \quad (7)$$

$$IR = ET_C - Pe_{\text{eff}} \text{ mm year}^{-1} \quad (8)$$

$$ET_{\text{green}} = ET_C - IR \text{ mm year}^{-1} \quad (9)$$

$$ET_C = K_c * ET_O \text{ mm year}^{-1} \quad (10)$$

$$ET_C = ET_C * \text{month in each year} \text{ mm year}^{-1} \quad (11)$$

$$WF_{\text{gray}} = ((\alpha * AR) / (C_{\text{max}} - C_{\text{nat}})) / Y \text{ m}^3 \text{ t}^{-1} \quad (12)$$

Green and blue water evapotranspiration of crops

The water demand of grapevine and olive crops was determined considering the crop water requirement (CWR) using the CROPWAT 8.0 software^{73,74}. This software allows the calculation of water supply for different crop patterns, amount of irrigation according to their typology, and crop yield, both under irrigated and non-irrigated conditions⁷⁵.

The crop evapotranspiration (ET_c) was calculated based on irrigation requirement (IR), assuming that water losses due to irrigation remain and return to the basin.

Where ET_c (mm/dec) is adjusted to decadal values (dec), in relation to irrigation efficiency throughout the crop growing season, under ideal growing conditions, it is assumed that the water requirements of the plantations are met⁷⁶, so that the sum of ET_{green} and ET_{blue} equals ET_c.

ET_{blue} was estimated according to Eq. (7), an equation that included the determination of I.R. (Eq. (8)). the program calculated Effective Precipitation (Peff) according to the USDA S.C. method. This parameter refers to the fraction of the total precipitation used to meet the irrigation needs of the crop. If (Peff) is higher than the total evapotranspiration of the crop, ET_{blue} is equal to zero indicating that no additional water is needed.

The value of ET_{green} was determined using Eq. (9), which considers the ET_c values calculated in Eqs. (10) and (11). The crop coefficient, K_c, considers the crop characteristics and the average effects of soil evaporation. The reference evapotranspiration, E_{T0} (mm month⁻¹), was calculated using the Penman–Monteith method with the CROPWAT 8.0 program. Climate data based on the latitude and analyzed period was also considered in the calculation²⁹.

Calculation of the gray water footprint of crops (WF_{gray})

The WF_{gray} is defined as the amount of water necessary to assimilate the residues and dilute them so that their concentration remains within the quality ranges, according to the regulations³⁸. WF_{gray} was estimated according to Eq. (12), where AR corresponds to the applied amount of fertilizer, α represents the leaching and runoff fraction of the product expressed as a percentage, C_{max} is the maximum acceptable concentration defined by quality standards, C_{nat} is the natural concentration of the pollutant, Y is the yield of agricultural production, and the agent used corresponds to nitrogen for each agricultural crop (AR)⁷².

The leaching fraction of this agent was 10%, and the C_{max} was 0.03 kg m³, based on NOM-001-SEMARNAT Mexican Emission Standard. The C_{nat} was set at 0.00001 kg m³ through the best-case scenario assessment according to the methodology of Hoekstra⁷².

Statistical analyses

Correlation analyses were performed for the hydrological variables according to the Pearson coefficient using STATISTICA 8.0. In addition, for the WF_{agricultural} ANOVA tests were computed.

Data availability

This entire data set, or specific portions of it, supporting the findings of this study are available from the author V.N. on request.

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References

- Li, Y., Gong, Y. & Huang, C. Construction of combined drought index based on bivariate joint distribution. *Alex. Eng. J.* **60**, 2825–2833. <https://doi.org/10.1016/j.aej.2021.01.006> (2021).
- Drisya, J., Kumar, S. & D., Evaluation of the drought management measures in a semi-arid agricultural watershed. *Environ. Dev. Sustain.* **25**(1), 811–833. <https://doi.org/10.1007/s10668-021-02079-4> (2023).
- López-Lambraño, A. A. *et al.* Supply and demand analysis of water resources. Case study: Irrigation water demand in a semi-arid zone in Mexico. *Agriculture* **10**(8), 333. <https://doi.org/10.3390/agriculture10080333> (2020).
- Zhang, Y., Fu, B., Feng, X. & Pan, N. Response of ecohydrological variables to meteorological drought under climate change. *Remote Sens.* **14**(8), 1920. <https://doi.org/10.3390/rs14081920> (2022).
- Orimoloye, I. R., Belle, J. A., Orimoloye, Y. M., Olusola, A. O. & Olofade, O. O. Drought: A Common Environmental Disaster. *Atmos* **13**(1), 111. <https://doi.org/10.3390/atmos13010111> (2022).
- Cook, B. I. *et al.* Megadroughts in the Common Era and the Anthropocene. *Nat. Rev. Earth Environ.* **3**, 741–757. <https://doi.org/10.1038/s43017-022-00329-1> (2022).
- van Loon, A. F. Hydrological drought explained. *Wiley Interdiscip. Rev. Water.* **2**, 359–392. <https://doi.org/10.1002/wat2.1085> (2015).
- Huang, Z., Yuan, X. & Liu, X. The key drivers for the changes in global water scarcity: Water withdrawal versus water availability. *J. Hydrol.* **601**, 126658. <https://doi.org/10.1016/j.jhydrol.2021.126658> (2021).
- Lian, X. *et al.* Multifaceted characteristics of dryland aridity changes in a warming world. *Nat. Rev. Earth Environ.* **2**, 232–250. <https://doi.org/10.1038/s43017-021-00144-0> (2021).
- Del Toro-Guerrero, F. & Kretzschmar, T. Precipitation-temperature variability and drought episodes in northwest Baja California. *México. J. Hydrol. Reg. Stud.* <https://doi.org/10.1016/j.ejrh.2019.100653> (2020).
- Bo, Y. *et al.* Additional surface-water deficit to meet global universal water accessibility by 2030. *J. Clean. Prod.* <https://doi.org/10.1016/j.jclepro.2021.128829> (2021).
- Chen, W., Kang, J. N. & Han, M. S. Global environmental inequality: Evidence from embodied land and virtual water trade. *Sci. Total Environ.* **783**, 146992. <https://doi.org/10.1016/j.scitotenv.2021.146992> (2021).
- Taylor, C. A. & Rising, J. Tipping point dynamics in global land use. *Environ. Res. Lett.* **16**(12), 125012. <https://doi.org/10.1088/1748-9326/ac3c6d> (2021).
- de Jager, A., Corbane, C. & Szabo, F. Recent developments in some long-term drought drivers. *Climate* **10**(3), 31. <https://doi.org/10.3390/cli10030031> (2022).

15. Yang, L. & Cormican, K. The crossovers and connectivity between systems engineering and the sustainable development goals: A scoping study. *Sustainability* **13**(6), 3176. <https://doi.org/10.3390/su13063176> (2021).
16. Yu, J. *et al.* Evaluating sustainable intensification levels of dryland agriculture: A focus on Xinjiang. *China. Ecol. Indic.* **158**, 111448. <https://doi.org/10.1016/j.ecolind.2023.111448> (2024).
17. Boonwichai, S., Shrestha, S., Babel, M. S., Weesakul, S. & Datta, A. Evaluation of climate change impacts and adaptation strategies on rainfed rice production in Songkhram River Basin, Thailand.. *Sci. Total Environ.* **620**, 189–201. <https://doi.org/10.1016/j.scitotenv.2018.10.201> (2019).
18. Albert, J. S. *et al.* Scientists' warning to humanity on the freshwater biodiversity crisis. *Ambio* **50**, 85–94. <https://doi.org/10.1007/s13280-020-01318-8> (2021).
19. Wang, Q., Huang, K., Liu, H. & Yu, Y. Factors affecting crop production water footprint: A review and meta-analysis. *Sustain. Prod. Consum.* **36**, 207–216. <https://doi.org/10.1016/j.spc.2023.01.008> (2023).
20. Hoekstra, A. Y. Green-blue water accounting in a soil water balance. *Adv. water Resour.* **129**, 112–117. <https://doi.org/10.1016/j.advwatres.2019.05.012> (2019).
21. Zhuo, L., Hoekstra, A. Y., Wu, P. & Zhao, X. Monthly blue water footprint caps in a river basin to achieve sustainable water consumption: The role of reservoirs. *Sci. Total Environ.* **650**, 891–899. <https://doi.org/10.1016/j.scitotenv.2018.09.090> (2019).
22. Rosa, L., Chiarelli, D. D., Rulli, M. C., Dell'Angelo, J. & D'Odorico, P. Global agricultural economic water scarcity. *Sci. Adv.* **6**(18), 6031. <https://doi.org/10.1126/sciadv.aaz6031> (2020).
23. Zisopoulou, K. & Panagoulia, D. An in-depth analysis of physical blue and green water scarcity in agriculture in terms of causes and events and perceived amenability to economic interpretation. *Water* **13**(12), 1693. <https://doi.org/10.3390/w13121693> (2021).
24. Falkenmark, M. & Rockström, J. The new blue and green water paradigm: Breaking new ground for water resources planning and management. *J. Water Resour. Plan. Manag.* **132**(3), 129–132. [https://doi.org/10.1061/\(ASCE\)0733-9496\(2006\)132:3\(129\)](https://doi.org/10.1061/(ASCE)0733-9496(2006)132:3(129)) (2006).
25. Marston, L. & Konar, M. Drought impacts to water footprints and virtual water transfers of the Central Valley of California. *Water Resour. Res.* **53**(7), 5756–5773. <https://doi.org/10.1002/2016WR020251> (2017).
26. Cao, X., Bao, Y., Li, Y., Li, J. & Wu, M. Unravelling the effects of crop blue, green and grey virtual water flows on regional agricultural water footprint and scarcity. *Agric. Water Manag.* **278**, 108165. <https://doi.org/10.1016/j.agwat.2023.108165> (2023).
27. Ponce-Oliva, R. D., Montevechio, E. A., Jorquera, F. F., Vásquez-Lavin, F. & Stehr, A. Water use and climate stressors in a multiuser river basin setting: Who benefits from adaptation?. *Water Resour. Manag.* **35**, 897–915. <https://doi.org/10.1007/s11269-020-02753-8> (2021).
28. Naorem, A. *et al.* Soil constraints in an arid environment—challenges, prospects, and implications. *Agronomy* **13**(1), 220. <https://doi.org/10.3390/agronomy13010220> (2023).
29. Sun, J. X. *et al.* Review on research status of virtual water: The perspective of accounting methods, impact assessment and limitations. *Agric. Water Manag.* **243**, 106407. <https://doi.org/10.1016/j.agwat.2020.106407> (2021).
30. Mekonnen, M. M. & Hoekstra, A. Y. Sustainability of the blue water footprint of crops. *Adv. Water Resour.* **143**, 103679. <https://doi.org/10.1016/j.advwatres.2020.103679> (2020).
31. Bisbis, M. B., Gruda, N. & Blanke, M. Potential impacts of climate change on vegetable production and product quality—A review. *J. Clean. Prod.* **170**, 1602–1620. <https://doi.org/10.1016/j.jclepro.2017.09.224> (2018).
32. Novoa, V. *et al.* Water footprint and virtual water flows from the Global South: Foundations for sustainable agriculture in periods of drought. *Sci. Total Environ.* **869**, 161526. <https://doi.org/10.1016/j.scitotenv.2023.161526> (2023).
33. Mian, H. R., Hu, G., Hewage, K., Rodriguez, M. J. & Sadiq, R. Drinking water quality assessment in distribution networks: A water footprint approach. *Sci. Total Environ.* **775**, 145844. <https://doi.org/10.1016/j.scitotenv.2021.145844> (2021).
34. Stahle, D. W. *et al.* Early 21st-century drought in Mexico. *Eos* **90**(11), 89–90. <https://doi.org/10.1029/2009EO110001> (2009).
35. Haeffner, M., Baggio, J. A. & Galvin, K. Investigating environmental migration and other rural drought-adaptation strategies in Baja California Sur, Mexico. *Reg. Environ. Change.* **18**, 1495–1507. <https://doi.org/10.1007/s10113-018-1281-2> (2018).
36. Martínez-Sifuentes, A. R. *et al.* Two centuries of drought history in the center of Chihuahua, Mexico. *Forests* **13**, 921. <https://doi.org/10.3390/f13060921> (2022).
37. Ivanova, A., Gámez, A. E. Plan estatal de acción ante el cambio climático para Baja California Sur La Paz, México (2012).
38. Manzanares, J. L. Retos para la producción vitivinícola en la región norte de Baja California: Una Caracterización hidrológica del valle de Guadalupe. *SAFER* **8**, 323164. <https://doi.org/10.7770/safer-VON0-art2085> (2020).
39. González, A. & S.. Cadena de valor económico del vino de Baja California, México. *Estud. Front.* **16**(32), 163–193 (2015).
40. Ibarra-García, M. V. & Talledos-Sánchez, E. Three cases of groundwater concentration in Mexico. *Water. Environ. J.* **15**, 35–44. <https://doi.org/10.17561/at.15.4649> (2020).
41. OEIDRUS. Estudio estadístico y geográfico sobre producción del Olivo en Baja California, Oficina Estatal de Información para el Desarrollo Rural Sustentable, México (2021).
42. Zhao, Z., Holbrook, N. J., Oliver, E. C., Ballester, D. & Vargas-Hernandez, J. M. Characteristic atmospheric states during mid-summer droughts over Central America and Mexico. *Clim. Dyn.* **55**, 681–701. <https://doi.org/10.1007/s00382-020-05283-6> (2020).
43. Zanin, P. R. & Satyamurty, P. Hydrological processes interconnecting the two largest watersheds of South America from multi-decadal to inter-annual time scales: A critical review. *Int. J. Climatol.* **40**(9), 4006–4038. <https://doi.org/10.1002/joc.6442> (2020).
44. Velasco, E. M., Gurdak, J. J., Dickinson, J. E., Ferré, T. P. A. & Corona, C. R. Interannual to multidecadal climate forcings on groundwater resources of the US West Coast. *J. Hydrol. Reg. Stud.* **11**, 250–265. <https://doi.org/10.1016/j.ejrh.2015.11.018> (2017).
45. Arriaga-Ramírez, S. & Cavazos, T. Regional trends of daily precipitation indices in northwest Mexico and southwest United States. *J. Geophys. Res.* <https://doi.org/10.1029/2009JD013248> (2010).
46. Macías-Carranza, V. & Cabello-Pasini, A. Climatología y evapotranspiración en valles vitivinícolas de Baja California. *Rev. Mexicana Cienc. Agric.* **12**(5), 849–863. <https://doi.org/10.29312/remexca.v12i5.2816> (2021).
47. Huntington, J. L. *et al.* Climate engine: Cloud computing and visualization of climate and remote sensing data for advanced natural resource monitoring and process understanding. *BAMS* **98**(11), 2397–2410. <https://doi.org/10.1175/BAMS-D-15-00324.1> (2017).
48. Mahlknecht, J., González-Bravo, R. & Loge, F. J. Water-energy-food security: A Nexus perspective of the current situation in Latin America and the Caribbean. *Energy* **194**, 116824. <https://doi.org/10.1016/j.energy.2019.116824> (2020).
49. Chitra-Tarak, R. *et al.* The roots of the drought: Hydrology and water uptake strategies mediate forest-wide demographic response to precipitation. *J. Ecol.* **106**(4), 1495–1507. <https://doi.org/10.1111/1365-2745.12925> (2018).
50. Pino, E., Montalván, I., Vera, A. & Ramos, L. La conductancia estomática y su relación con la temperatura foliar y humedad del suelo (Olea europaea L.), en el cultivo del olivo en periodo de maduración de frutos, en zonas áridas. La Yarada, Tacna, Perú. *Idesia* **37**(4), 55–64. <https://doi.org/10.4067/S0718-34292019000400055> (2021).
51. Salgado, J., Palacios, O., Galvis, A. & Mejía, E. Water quality effect on the Valle de Guadalupe aquifer in the agricultural soils salinity. *Rev. Mex. Cienc. Agríc.* **3**(1), 79–95 (2012).
52. Li, J., Lin, M. & Feng, Y. Improved grey water footprint model based on uncertainty analysis. *Sci. Rep.* **13**(1), 7100. <https://doi.org/10.1038/s41598-023-34328-z> (2023).
53. Azócar, G. *et al.* Climate change perception, vulnerability, and readiness: inter-country variability and emerging patterns in Latin America. *J. Environ. Stud. Sci.* **11**(1), 23–36. <https://doi.org/10.1007/s13412-020-00639-0> (2021).
54. Banwell, N., Gesche, A., Rojas, O. & Hostettler, S. Barriers to the implementation of international agreements on the ground: Climate change and resilience building in the Araucanía Region of Chile. *Int. J. Disaster Risk Reduct.* **50**, 1011703. <https://doi.org/10.1016/j.ijdrr.2020.101703> (2020).

55. Duan, Y., Wang, W., Zhuo, L., Liu, Y. & Wu, P. Regional blue and green water-saving potential and regulation paths for crop production: A case study in the Yellow River Basin. *Agric. Water Manag.* **291**, 108631. <https://doi.org/10.1016/j.agwat.2023.108631> (2024).
56. Davis, K. F., Rulli, M. C., Seveso, A. & D'Odorico, P. Increased food production and reduced water use through optimized crop distribution. *Nature Geosci.* **10**(12), 919–924. <https://doi.org/10.1038/s41561-017-0004-5> (2017).
57. Degefu, D. M. *et al.* Mapping monthly water scarcity in global transboundary basins at country-basin mesh based spatial resolution. *Sci. Rep.* **8**(1), 2144. <https://doi.org/10.1038/s41598-018-20032-w> (2018).
58. Vicente-Serrano, S. M., Quiring, S. M., Pena-Gallardo, M., Yuan, S. & Dominguez-Castro, F. A review of environmental droughts: Increased risk under global warming?. *Earth Sci Rev.* **201**, 102953. <https://doi.org/10.1016/j.earscirev.2019.102953> (2020).
59. Novoa, V. *et al.* Sustainability assessment of the agricultural water footprint in the Cachapual River basin. *Chile. Ecol. Indic.* **98**, 19–28. <https://doi.org/10.1016/j.ecolind.2018.10.048> (2019).
60. Novoa, V. *et al.* Understanding agricultural water footprint variability to improve water management in Chile. *Sci. Total Environ.* **670**, 199–199. <https://doi.org/10.1016/j.scitotenv.2019.03.127> (2019).
61. Barbier, E. B. Long run agricultural land expansion, booms and busts. *Land Use Policy.* **93**, 103808. <https://doi.org/10.1016/j.LANDUSEPOL.2019.01.011> (2020).
62. Gleeson, T. *et al.* Illuminating water cycle modifications and Earth system resilience in the Anthropocene. *Water Resour. Res.* <https://doi.org/10.1029/2019WR024957> (2020).
63. Cassman, K. G. & Dobermann, A. Nitrogen and the future of agriculture: 20 years on. *Ambio.* **51**(1), 17–24. <https://doi.org/10.1007/s13280-021-01526-w> (2022).
64. IIEG (Instituto de Información Estadística y Geográfica de Jalisco). *Producto Interno Bruto por entidad federativa a precios constantes, Producto Interno Bruto.* https://iieg.gob.mx/ns/?page_id=1162. (Accessed 01 August 2023) (2019).
65. Cabello-Pasini, A., Macias-Carranza, V. & Mejía-Trejo, A. Efecto del mesoclima en la maduración de uva nebbiolo (*Vitis vinifera*) en el valle de Guadalupe, Baja California, México. *Agrociencia* **51**(6), 617–633 (2017).
66. ESRI (Environmental Systems Research Institute). *ArcGIS 9.3* (ESRI, 2009).
67. POEBC. *2016 Programa Ambiental Estratégico de la Región Vitivinícola de Valle de Guadalupe en el Municipio de Ensenada, Baja California, Secretaría de Protección al Ambiente de Baja California* (2016).
68. Cavazos, T. & Rivas, D. Variability of extreme precipitation events in Tijuana. *Mexico. Clim. Res.* **25**(3), 229–243. <https://doi.org/10.3354/cr025229> (2004).
69. Colorado-Ruiz, G. & Cavazos, T. Trends of daily extreme and non-extreme rainfall indices and intercomparison with different gridded data sets over Mexico and the southern United States. *Int. J. Climatol.* **41**(11), 5406–5430. <https://doi.org/10.1002/joc.7225> (2021).
70. Searcy, J. K. & Hardison, C. H. *Double-Mass Curves (No. 1541)* (US Government Printing Office, 1960).
71. Valiente, Ó. M. Sequía: definiciones, tipologías y métodos de cuantificación. *Investig. Geogr.* **26**, 59–80 (2001).
72. Hoekstra, A. Y., Chapagain, A. K., Aldaya, M. M. & Mekonnen, M. M. *2011 The Water Footprint Assessment Manual: Setting the Global Standard*, London, UK: EarthScan (Routledge, 2011).
73. Smith, M. *CROPWAT: A computer program for irrigation planning and management* (Food & Agriculture Org, 1992).
74. FAO. *Software CROPWAT 8 standard crop and soil data.* <http://www.fao.org/land-water/databases-and-software/cropwat/en/> (Accessed 12 August 2023). (2020).
75. Swennenhuis, J. *CROPWAT 8.0* (Water Resources Development and Management Service of FAO, 2009).
76. Gabr, M. E. Modelling net irrigation water requirements using FAO-CROPWAT 8.0 and CLIMWAT 2.0: A case study of Tina Plain and East South ElKantara regions North Sinai Egypt. *Arch. Agron. Soil Sci.* **68**(10), 1322–1337. <https://doi.org/10.1080/03650340.2021.1892650> (2022).

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Author contributions

V.N. and C.R. conceptualization, writing, methodology, results, graphics, discussion. O.R. data analysis, writing, revision, editing, visualization, supervision, project management. R.A. gray component analysis, writing, revision, editing. R.M.S cartographies, land use, drafting, revision, editing.

Competing interests

The authors declare no competing interests.

Additional information

Correspondence and requests for materials should be addressed to R.A.-R.

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